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SPATIO-TEMPORAL VARIABILITY OF ANNUAL PM_{2.5} CONCENTRATIONS AND POPULATION EXPOSURE ASSESSMENT IN SERBIA FOR THE PERIOD 2001–2016

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Abstract: The long-term exposure to PM_{2.5} (particulate matter with aerodynamic diameter $\leq 2.5 \mu\text{m}$) is the leading global health risk factor. The spatio-temporal variability of annual values of the total PM_{2.5} concentrations in Serbia is analyzed using the high-resolution ($0.01^\circ \times 0.01^\circ$) gridded data set V4.EU.02 for the period 2001–2016. Population counts and density data were used to calculate the population exposure while the urban land cover data were used to estimate the relations between the size of urban area and the concentration of PM_{2.5}. The mean annual values vary in the range $13.93 \mu\text{g}/\text{m}^3$ – $28.91 \mu\text{g}/\text{m}^3$. The regional differences exist, but the highest values were obtained for urban environments ($> 22.5 \mu\text{g}/\text{m}^3$). Negative trend of annual PM_{2.5} is present in most of the parts of the Serbian territory, especially in the eastern parts reaching $-0.37 \mu\text{g}/\text{m}^3$ per year ($p \leq .05$). More than 99% of the territory of Serbia has the mean annual values under the national safe limit established by regulation ($25 \mu\text{g}/\text{m}^3$), but comparing to the World Health Organization (WHO) guideline ($10 \mu\text{g}/\text{m}^3$), all the territory is above the safe level. In line with the trend of urbanization, there is a clear upward trend in the number of population exposed to the higher concentrations of PM_{2.5}. The share of the population exposed to values higher than $25 \mu\text{g}/\text{m}^3$ increased from 6.65% in 2005 to 11.40% in 2015, while comparing to WHO standard, the total population in Serbia is exposed to the values higher than the safe one.

Keywords: air pollution; PM_{2.5}; trend; population exposure; Serbia

Introduction

Atmospheric aerosol particles or particulate matter (PM) are classified as air pollutants that are directly emitted (primary PM) or formed in the atmosphere from precursor gases through reactions (secondary PM). They can either have natural (marine aerosol, mineral dust, biological aerosol, volcanic

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ash, etc.) or anthropogenic origin (fuel combustion for power generation, domestic heating and transport, industry and waste incineration, etc.) (European Environment Agency [EEA], 2018; Fuzzi et al., 2015; Tasić, Rajšić, Novaković, & Mijić, 2006). The PM consists of complex mixtures of organic and inorganic compounds, solid and liquid particles, with a wide spectrum of physical and chemical properties and different aerodynamic diameter (*ultrafine* with diameter $\leq 0.1 \mu\text{m}$ —PM _{≤ 0.1} , *fine* with diameter $> 0.1 \leq 2.5 \mu\text{m}$ —PM _{$> 0.1 \leq 2.5$} , and *coarse* with diameter $> 2.5 \leq 10 \mu\text{m}$ —PM _{$> 2.5 \leq 10$}) that determine their climate and health-related effects (Reizer & Juda-Rezler, 2016; Vicente & Alves, 2018).

The air pollution has been recognized as a risk factor for human health by the World Health Organization (WHO) since 1958 (WHO Regional Office for Europe, 2017). Ambient (outdoor) air pollution was estimated to cause 4.2 million premature deaths worldwide in 2016, mainly due to PM_{2.5} exposure (WHO, 2018). Numerous studies found a strong exposure-response to the airborne PM in the form of short-term effects (hospital admissions) and long-term or cumulative health effects (morbidity, lung cancer, cardiovascular and cardiopulmonary diseases, etc.) (Beelen et al., 2008; Brauer et al., 2003; Brunekreef & Holgate, 2002; Pope & Dockery, 2006; Vodonos, Awad, & Schwartz, 2018). The PM_{2.5} are small in size and rich in toxic substances (known as lung particles) which determine their greater impact on human health than the coarse PM (Shisong et al., 2018). Deaths worldwide attributable to ambient PM_{2.5} increased from 3.5 million in 1990 to 4.2 million in 2015, making long-term exposure to ambient PM as a fifth-ranked global health risk factor in 2015 (Cohen et al., 2017).

According to the EEA (2018), the PM_{2.5} concentrations reached in 2015 were responsible for about 422,000 premature deaths emerging from long-term exposure in Europe (over 41 countries). The highest numbers of premature deaths and YLL (Years of life lost) were estimated for the countries with the largest populations (Germany, Italy, Poland, France, and the United Kingdom). When considering YLL per 100,000 inhabitants, the largest impact was recorded in central and eastern European countries (Bulgaria, Serbia, North Macedonia, and Hungary) where the highest PM_{2.5} concentrations were also registered. The urban population is mostly affected by air pollution due to numerous activities (traffic, industry, household heating, etc.) that increase the level of PM in the atmosphere (Karagulian et al., 2015). Yang et al. (2018) stated that the increased PM_{2.5} concentration in cities became the major social issue together with rapid urbanization in many regions of the world. Taking into account the effects of PM on the humans, especially when it comes to PM_{2.5}, Shaddick et al. (2018) emphasize the need for assessing the population exposure at the national and lower territorial units' level.

Previous studies about air pollution by PM in Serbia are mainly focused on PM₁₀, while those related to PM_{2.5} are rare and often include PM₁₀, PM_{0.1} or other air pollutants (NO₂, SO₂, etc.). Some authors discussed the chemical composition and physical properties of airborne particles, related to different sources of air pollution (e.g., Cvetković, Logar, Rosić, & Čirić, 2012; Joksić et al., 2009; Šerbula, Antonijević, Milošević, Milić, & Ilić, 2010), while the conducted epidemiological studies examined the association of exposure to airborne PM with health effects (Kovačević, 2019; Stevanović, Jovašević-Stojanović, & Jović-Stošić, 2016; WHO Regional Office for Europe, 2019). Todorović, Radenković, Rajšić, and Ignjatović (2019) point out that the health risk assessment studies for Serbia are scarce due to the lack of long-term air quality monitoring data and medical records, such as hospital admissions and deaths due to specific causes, on daily basis. In order to overcome the disadvantage associated with insufficient and poor spatial coverage of measurement sites (Vuković et al., 2015), several studies use the method of biomonitoring to assess air quality (Aničić, Spasić, Tomašević, Rajšić, & Tasić, 2011; Stamenković et al., 2016). Certain studies discuss low-cost sensing technologies for air quality monitoring and exposure assessment (Jovašević-Stojanović et al., 2015; Tasić, 2017). The studies are mainly related to the so-called "hot spots", larger cities, especially for (the capital) Belgrade (Rajšić,

Tasić, Novaković, & Tomašević, 2004; Tasic, Rajsic, Novakovic, & Mijic, 2007) and industrial centers, e.g. the City of Bor (Djordjevic et al., 2013; Tasić, Milošević, Kovačević, Jovašević-Stojanović, Dimitrijević, 2012). On the regional level, Malinović-Miličević, Mihailović, Nikolić-Đorić, and Jevtić (2015) analyzed the impact of air pollution sources (household heating and traffic) on air quality in Vojvodina region for the period 2001–2008. So far, there is a lack of studies on the spatio-temporal variability of PM_{2.5} for the whole territory of Serbia as well as on population exposure.

Assessing the effects of PM_{2.5} exposure is often limited by the absence of ground-level measurements with convenient spatial and temporal coverage. In recent years, remote sensing data have enabled the estimation of the global distribution of ambient PM_{2.5} concentrations. Satellite data are processed together with Chemical Transport Model (CTM) simulations and ground-level measurements to produce the high-resolution datasets enabling an insight into long-term (decade or more) spatio-temporal changes and offer information about the magnitude of human exposure for epidemiological studies (van Donkelaar et al., 2010; van Donkelaar, Martin, Brauer, & Boys, 2015; van Donkelaar et al., 2016; van Donkelaar, Martin, Li, & Burnett, 2019). These data are used in studies on global (Li, Han, Jin, Zhang, & Wang, 2019; Wang et al. 2019; Yang et al., 2018), regional and national levels (Bai, Ma, Chang, & Gao, 2019; Larkin, van Donkelaar, Geddes, Martin, & Hystad, 2016; Lu, Mao, Yang, Zhao, & Xu, 2018). In this research, spatial variability and temporal trend are analyzed using the high-resolution (0.01° × 0.01°) dataset of the total annual PM_{2.5} concentrations in Serbia for the period 2001–2016. In addition, the variability of PM_{2.5} exposure is analyzed depending on the size of the urban area. The same spatial resolution datasets about population count and density are used to estimate the population exposure.

Study area

Serbia is positioned in southeast Europe on the Balkan Peninsula and covers an area of 88,499 km² (Figure 1) (Statistical Office of the Republic of Serbia, 2019). The northern part of the country is represented by low land, which is the southeast part of the Pannonian plain. The altitude increases to the south where the hills and mountains interspersed by river valleys and basins are dominant forms of relief. The highest altitude is 2,656 m a.s.l. (Đeravica Peak, Prokletije Mountain) situated on the southwest of the country.

According to the last Census data from 2011 (Statistical Office of the Republic of Serbia, 2014), the number of inhabitants is 7,186,862

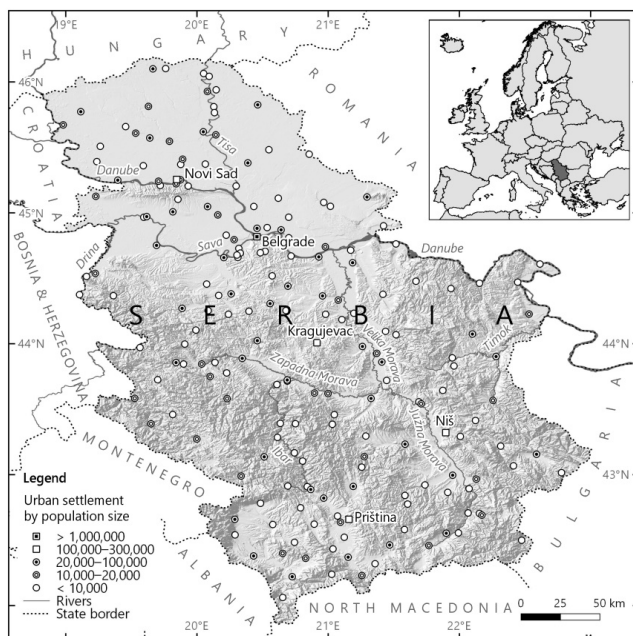


Figure 1. Study area.

(without Kosovo and Metohija¹). The population is mainly concentrated at lower altitudes in northern and central areas and along the river valleys, such as valleys of Danube, Sava, and Velika Morava rivers. The majority of the population (~60%) lives in urban settlements and the largest number of them is Belgrade, with the population of ~ 1.7 million in the total city area, i.e., ~ 1.2 million in urban city area (Drobnjaković & Spalević, 2017).

Data and methods

The European Regional Estimates (V4.EU.02) dataset (Atmospheric Composition Analysis Group, n.d.) is used to calculate the mean annual values of the total PM_{2.5} (µg/m³) for the territory of Serbia in the period 2001–2016. It presents the combined geophysical-statistical estimation of PM_{2.5} over Europe using available measurements obtained from the EEA Air Quality e-Reporting system. The estimation of ground-level fine particulate matter (PM_{2.5}) in total (µg/m³) is based on combining Aerosol Optical Depth (AOD) retrievals from the NASA MODIS, MISR, and SeaWiFS instruments with the GEOS-Chem chemical transport model, and subsequently calibrated to regional ground-based observations using Geographically Weighted Regression (GWR) (van Donkelaar et al., 2019). The spatial resolution is 0.01° × 0.01°. Besides the regional aspect of variability, this study explores the changes in PM_{2.5} for different altitudes. Using the European Digital Elevation Model (EU-DEM), version 1.1 (EEA, 2013), several hypsometric zones for the territory of Serbia were singled out (< 200 m, ≥ 200 < 500 m, ≥ 500 < 1000 m, ≥ 1000 < 1500 m, and ≥ 1500 m a.s.l.) and the average mean annual PM_{2.5} is calculated. Digital elevation model has been resampled to fit the grid of PM_{2.5} data to perform a zonal grid statistics report. In order to explore the temporal variability of annual PM_{2.5} values, the linear trend is calculated for the observed period, while for determining the significance of a trend, the Student's t-test is used.

Many urban activities contribute to the increase of PM concentration. This study aims to analyze the variation of PM_{2.5} concentration depending on urban surface size for the territory of Serbia. The Global Human Built-up And Settlement Extent (HBASE) Dataset From Landsat, v1 (2010) (Wang, Huang, Brown de Colstoun, Tilton, & Tan, 2017) is used to obtain the data on the urban extent and “man-made” surfaces with a spatial resolution of 30 m in 2010 as the reference year. For the purpose of this study, urban land cover, i.e., continual built-up surfaces, are categorized into several classes: < 5 km², ≥ 5 < 10 km², ≥ 10 < 20 km², ≥ 20 < 50 km², ≥ 50 < 100 km², and ≥ 100 km². For each class, the average annual PM_{2.5} is calculated for the target year (2010).

The estimation of the population exposure to ambient air pollution is the first step in the impact assessment studies. In order to calculate the population exposure to PM_{2.5}, two different datasets were used: UN WPP-Adjusted Population Count, v4.11 (Center for International Earth Science Information Network [CIESIN], 2018a) and UN WPP-Adjusted Population Density, v4.11 (CIESIN, 2018b). The data are available as gridded datasets with spatial resolution of 30 arc-second. The analyses were conducted for 2005, 2010, and 2015 and population count and density were calculated for annual values of PM_{2.5}. All the calculations in the study were performed using Quantum Geographical Information System (QGIS) (version 2.18.4) and the System for Automated Geoscientific Analyses Geographical Information System (SAGA GIS) (version 4.0.1, Conrad et al., 2015).

¹According to Kokotović Kanazir et al. (2017), the estimated number of inhabitants (based on United Nations statistics) in the Republic of Serbia (including Kosovo and Metohia) was 9,854,000 in 2011.

Results and discussion

The variability in the shares of sources that contribute to the emission of PM_{2.5} results in various spatio-temporal patterns. The mean annual values of PM_{2.5} total (2001–2016) for the territory of Serbia vary in the range 13.93 µg/m³–28.91 µg/m³. The calculated values are classified into several categories (together with the percentage of territory with these values): < 15 µg/m³ (3.27%), ≥ 15 < 17.5 µg/m³ (20.27%), ≥ 17.5 < 20 µg/m³ (38.41%), ≥ 20 < 22.5 µg/m³ (36.36%), ≥ 22.5 < 25 µg/m³ (1.01%), ≥ 25 < 27.5 µg/m³ (0.64%), and ≥ 27.5 µg/m³ (0.04%) (Figure 2). The most prominent feature are the isolated “islands” of the highest values of PM_{2.5} (> 22.5 µg/m³) that are spatially overlapped with the largest urban settlements on the territory of Serbia. Generally, higher values are obtained for the northern part of Serbia (20–22.5 µg/m³). Central and southern areas show values of 17.5–20.0 µg/m³, while the lowest ones are in the southeast (15–17.5 µg/m³) and southwest (< 15 µg/m³) parts of the country. The analysis of the hypsometric zones and mean annual PM_{2.5} (2001–2016) show a general tendency that, with an increase of altitude, the mean annual PM_{2.5} decreases (Table 1). The average mean annual PM_{2.5} has values of 21.08 µg/m³ and 16.92 µg/m³ for the terrain of < 200 m and ≥ 1,500 m, respectively. The minimum and maximum values are similar within the zones with altitudes up to 1,000 m a.s.l., but with the higher difference between them comparing

to the terrain with altitude more than 1,000 m a.s.l. This can be explained by the presence of “hot spots” of air pollution, i.e., urban settlements that are mainly situated in lower altitudes. The percentage of the territory of hypsometric zones with specified mean annual PM_{2.5} values is presented on Figure 3. It is shown that the higher percentage of the territory in lower altitude zones is exposed to the higher values of PM_{2.5}. This pattern follows the spatial distribution of the population in Serbia.

The mean annual PM_{2.5} concentration limit established by the European Parliament and the Council of the European Union (2008) is 25 µg/m³. This is significantly higher than the level considered as safe by WHO, i.e., specified as the guideline value of 10 µg/m³ (WHO, 2006). The legislation in Serbia is in line with Directive 2008/50/EC and the defined limit for annual PM_{2.5} (Stage 1) is 25 µg/m³ (Uredba o uslovima za monitoring i zahtevima kvaliteta vazduha, 11/2010,

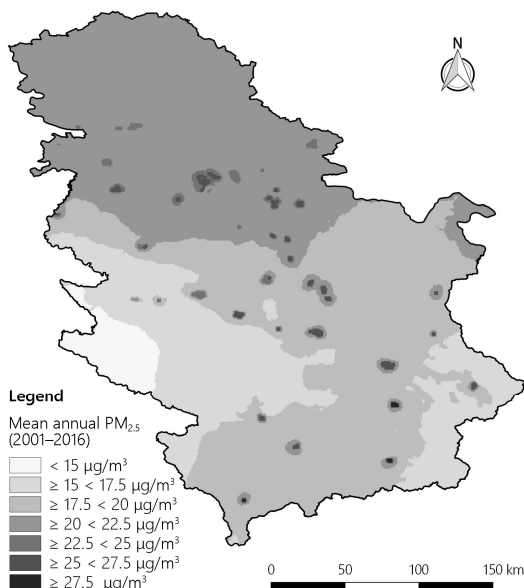


Figure 2. The mean annual values of PM_{2.5} (µg/m³) in Serbia for period 2001–2016.

Table 1

The average (*M*), standard deviation (*SD*), maximum (*min*), and minimum (*max*) values of mean annual PM_{2.5} (µg/m³) for hypsometric zones in Serbia

Hypsometric zones	<i>M</i>	<i>SD</i>	<i>min</i>	<i>max</i>
< 200 m	21.08	0.95	15.88	27.38
≥ 200 m < 500 m	18.90	1.32	14.49	28.91
≥ 500 m < 1000 m	17.51	1.26	14.04	27.40
≥ 1000 m < 1500 m	16.21	1.20	13.93	19.30
≥ 1500 m	16.92	0.78	14.49	18.38

75/2010, 63/2013). According to the results, it can be concluded that more than 99% of the territory of Serbia has mean annual values under this limit, but comparing to WHO standard, the whole territory is above the safe level. Even comparing to the value of 12 µg/m³ imposed by the U.S. Environmental Protection Agency for USA (Giannadaki, Giannakis, Pozzer, & Lelieveld, 2018), the whole territory of Serbia has higher mean annual PM_{2.5} concentrations.

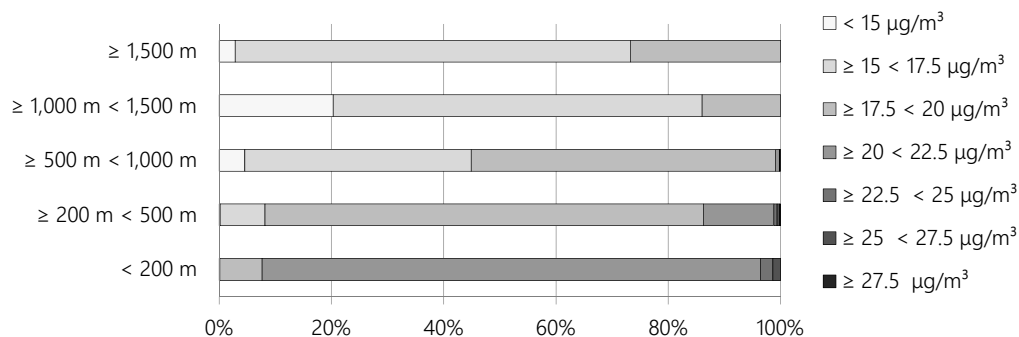


Figure 3. The percentage (%) of the territory of hypsometric zones with specified mean annual values of PM_{2.5} (µg/m³).

Multiple factors influence the level of PM_{2.5} concentrations in the atmosphere. Kiesewetter and Amann (2014) stated the importance of long-range transport of PM and precursor gases, from Europe-wide transboundary transport to local emissions within an urban area. Also, different source sectors that contribute to emissions vary for different territories. In Serbia, the largest contributors to the emission of gases (precursors of secondary PM) in 2016 were energy production and distribution (49.6% of NO_x and 93% of SO₂) and road traffic (24.3% of NO_x), as well as agriculture (84.7% of NH₃). The district heating plants with a capacity below 50 MW and individual domestic household heating contribute with 21.68% of non-methane volatile organic compounds (NMVOCs) (Lekić, 2018). The contribution of fossil fuels dominates (88%) in primary energy consumption in Serbia (Krunić Lazić, 2018). According to Jovanović (2019), district heating plants with a capacity below 50 MW and individual household heating contribute with 75% in the total emissions of PM_{2.5} in Serbia. In the dwellings without district and central heating (57.2% of the total occupied dwellings in Serbia), wood is mostly used for heating (63.3%), while electricity (15.2%) and coal (13.4%) are used less. In dwellings with central heating (20.6% of the total occupied dwellings in Serbia), wood (46.2%) is predominantly used for heating, followed by coal (25.4%) and gaseous fuels (18.4%), while in dwellings with district heating (22.1% of the total occupied dwellings in Serbia), gaseous fuels (30.4%) and oil fuels (26.1%) are mostly used, followed by coal and wood (21.7% and 17.3%) (Statistical Office of the Republic of Serbia, 2013). The household heating is the leading source of the emissions of PM_{2.5} in many regions in the world (Karagulian et al., 2015; Saarikoski et al., 2008; Samek et al. 2017; Xiao, Ma, Li, & Liu, 2015). The road transport sector contributed with 6%, while the share of industry was 9% in the total PM_{2.5} emissions in 2018 in Serbia (Jovanović, 2019). The relatively small share of industry in the emissions can be explained by the sharp decline in industrial production since the 1990s in Serbia (Hadžić & Zeković, 2019), but on the local level, in the larger industrial centers, it significantly contributes to PM emissions (PM₁₀ and PM_{2.5}) (e.g., Stevanović, Jovašević-Stojanović, & Jović Stošić, 2016).

The analysis of temporal variability showed a negative trend of annual PM_{2.5} for most of the Serbian territory in the period 2001–2016 (Figure 4). The highest statistically significant negative trend values are obtained for eastern and western parts of the country (in some areas in the east they reach up to $-0.37 \mu\text{g}/\text{m}^3$ per year, $p \leq .05$). The positive trend values are obtained only for small areas in the south that match urban settlements and reach up to $0.19 \mu\text{g}/\text{m}^3$ per year. These results are in accordance with the findings of studies exploring the spatial distribution and the trends of annual PM_{2.5} using similar datasets and periods. Using satellite data (aerosol optical depth), CTM and ground-measurements, Boys et al. (2014) produced the high-resolution ($1^\circ \times 1^\circ$) global dataset of PM_{2.5} for the period 1998–2012 and singled out four regions with a statistically significant trend: negative trend in Eastern US, and positive in Arabian Peninsula, South Asia, and East Asia. For the region of Southeast Europe (Serbia), they observed the absence of a statistically significant trend or significant negative trend (in the range from -0.25 to $-0.50 \mu\text{g}/\text{m}^3$ per year, $\alpha = 1\%$ and $\alpha = 5\%$).

Yang et al. (2018) showed an increasing trend of PM_{2.5} along with the urbanization during 1998–2015, but with a variety of evolutionary relations in different countries and regions. They stated that most countries in East, Southeast Asia and South Asia, and some African countries show rapid increase in both urbanization and PM_{2.5} while a decrease of PM_{2.5} concentrations are found in most European and American countries. Van Donkelaar et al. (2019) produced the high-resolution ($1 \text{ km} \times 1 \text{ km}$) dataset of the chemical composition of PM_{2.5} for North America and found a decrease in population-weighted PM_{2.5} concentrations since 2000 which can be explained by a reduction in sulfate and organic matter. According to EEA (2019), concerning the period 1990–2017, in EEA-33 region, the emission of NO_x, SO_x, and NMVOCs decreased by 57%, 82%, and 54% (61%, 91%, and 58% in EU-28), respectively. The NH₃ emissions have decreased by 18% (23% in EU-28), but have been continuously increasing since 2014. It is also stated that the three largest sectors (Energy production and distribution, Energy use in industry, and Commercial, institutional and household) have reduced the emissions of SO_x by more than 75% since 1990, but in the EEA-33 region (Southeast Europe), the power and heat generation still remain the most significant sources of SO_x contributing to over a half of the current total emissions. Besides the reduction in Energy production and distribution sector, the high impact on NO_x has the decrease in the emissions in road transport sector. Agriculture dominates in the emissions of NH₃ (92% of total emission in EEA-33). The emissions of primary PM_{2.5} decreased by 28% mainly due to the reductions in electricity generation and heavy industrial sectors. The households are still the largest contributor of PM_{2.5} in Southeast Europe. In Serbia, the emissions

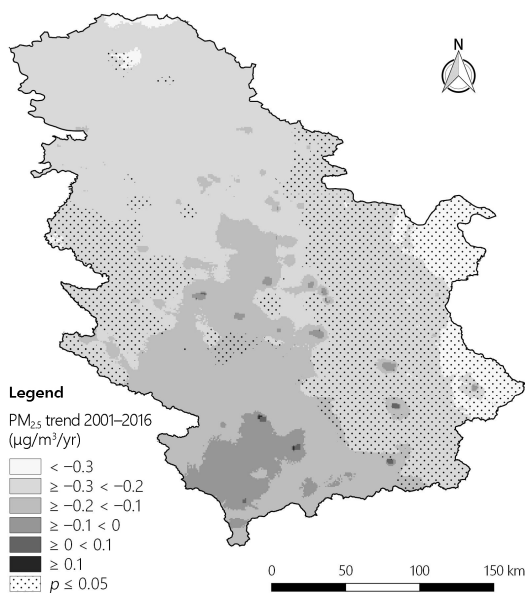


Figure 4. Trend of annual values of PM_{2.5} (µg/m³ per year) in Serbia for period 2001–2016.

Table 2

The average (M), standard deviation (SD), maximum (min) and minimum (max) of PM_{2.5} (µg/m³) for urban land cover surfaces for 2010 in Serbia

Classes	M	SD	min	max
< 5 km ²	21.12	1.67	15.20	26.00
≥ 5 km ² < 10 km ²	21.28	1.91	16.10	25.20
≥ 10 km ² < 20 km ²	22.22	2.11	17.90	28.50
≥ 20 km ² < 50 km ²	23.25	1.66	18.80	26.20
≥ 50 km ² < 100 km ²	23.11	1.59	20.00	26.40
≥ 100 km ²	23.33	0.97	22.00	25.30

of NO_x and SO_x showed a slow decrease, the emissions of NMVOCs showed a very slow decrease, while the emissions of NH₃ and PM₁₀ did not show any significant change in the period 1990–2016 (Lekić, 2018).

Both coarse and fine PM are present in urban environments. Many studies showed that their proportion varies substantially between cities depending on the local topography, meteorology and specific PM sources (Querol et al, 2008; Rogula-

Kozłowska et al., 2014; Perrone et al., 2018). Also, within one urban area, heterogeneous distribution patterns of PM_{2.5} can exist (Cheng & Wang-Li, 2019). The highest mean annual PM_{2.5} values are observed for urban settlements in Serbia. The analysis of PM_{2.5} values and urban land cover for 2010 showed that with an increase of the size of urban land cover, the annual PM_{2.5} increased (Table 2). It is similar for the minimum values, but the maximum values are found in the class ≥ 10 < 20 km² (value of 28.50 µg/m³). In many cities in Serbia, the problem of increased PM_{2.5} concentrations is related to district heating plants and a large share of individual household heating during the winter. This is especially expressed for urban settlements positioned in basins where local morphological and meteorological conditions contribute to the increased pollutant levels. The examples are the cities of Valjevo and Užice in western Serbia, where the observed records are the results of several factors during the winter season such as individual households heating, basin position, poor air ventilation, etc. In the city of Užice, in 2016, daily PM₁₀ reached a value of 438 µg/m³ (Radović & Jovanović, 2017), while in the city of Valjevo, in 2017, daily PM₁₀ reached a value of 806 µg/m³ (the upper limit defined by regulation is 50 µg/m³) (Radović & Jovanović, 2018). Unfortunately, the measurements of PM_{2.5} do not exist, so it is not possible to make the comparison with this record.

For assessing the impact of air pollution on humans, it is of the greatest importance to determine the population's exposures to the defined pollution levels. Population count and density data are related to the annual values of PM_{2.5} for three years: 2005, 2010, and 2015. The results are presented in Table 3. Certain differences between the observed years exist, but with a clear upward trend in the share of the population exposed to the higher values of PM_{2.5}. While the population density in the most polluted areas increases over time, the distribution of the population (in %) varies for the specified PM_{2.5} classes between years. The share of the population exposed to values higher than the nationally established annual safe limit of 25 µg/m³ increased from 6.65% in 2005 to 11.4% in 2015. Comparing to the WHO limit (10 µg/m³), the total population in Serbia is exposed to higher values during the observed years. In recent decades, urban settlements in Serbia have recorded population growth (Drobnjaković & Spalević, 2017). This process is accompanied with the increase in traffic activity, household heating, industry and other urban activities that increase air pollution in the urban environment. It is shown that the areas with the highest PM_{2.5} have the highest population density (in 2015, for areas with PM_{2.5} > 27.5 µg/m³, the average population density is 2,339.28 inhabitants per km²).

Table 3

The population counts (% of inhabitants) and average population density (inhabitants per km²) for specified values of annual PM_{2.5} (µg/m³) for 2005, 2010, and 2015 in Serbia

PM _{2.5} classes (µg/m ³)	2005		2010		2015	
	Population (%)	Population density (inh./km ²)	Population (%)	Population density (inh./km ²)	Population (%)	Population density (inh./km ²)
< 15	0.03	11.24	0.02	9.25	3.35	20.80
≥ 15 < 17.5	8.68	43.44	2.83	32.00	22.90	65.05
≥ 17.5 < 20	27.01	86.34	19.02	57.39	18.50	109.11
≥ 20 < 22.5	14.30	78.74	46.36	98.12	32.65	114.92
≥ 22.5 < 25	43.33	170.51	25.86	362.54	11.20	588.12
≥ 25 < 27.5	6.65	1453.08	5.88	1977.12	9.10	1391.12
≥ 27.5	0	0	0.03	284.13	2.30	2339.28

The air pollution problem is recognized as a global issue and numerous activities have been undertaken through many agendas (i.e., “Transforming our world: the 2030 Agenda for Sustainable Development” (United Nations, 2015)) and action programs (i.e., “Living well, within the limits of our planet” (European Parliament & Council of the European Union, 2013)). Van Donkelaar et al. (2016) found that the global population-weighted annual average PM_{2.5} concentration is 32.6 µg/m³ (1998–2015) and it is driven by high concentrations in several Asian and African regions. The results of this study showed that Serbia is below this worldwide average, but still in the health risk zone, especially concerning the WHO limit. The Report “Health impact of ambient air pollution in Serbia – A call to action” (WHO Regional Office for Europe, 2019) showed that nearly 3,600 premature deaths every year (2010–2015) are attributable to the exposure to PM_{2.5} in 11 cities in Serbia. It is evident that Serbia needs an adequate public policy in mitigating this health risk.

Conclusion

One of the biggest challenges facing the world's population is air pollution problem and finding the solution how to reduce and control the emissions of pollutants in order to prevent negative health effects. Long-term exposure to ambient PM_{2.5} becomes one of the leading global health risk factors. The satellite-derived data products help to overcome the insufficient number of ground-level measurements and give an insight in spatio-temporal variability of PM_{2.5}. These datasets provide valuable information for epidemiological studies and health impact assessment. This study analyzes the spatial variability and temporal evolution of total PM_{2.5} and estimates population exposure in Serbia for the period 2001–2016. The main findings are:

- The mean annual PM_{2.5} values (2001–2016) vary in the range 13.93 µg/m³–28.91 µg/m³. The higher values are obtained for the northern part of Serbia (20–22.5 µg/m³). Central and southern areas show values of 17.5–20.0 µg/m³, while the lowest values are recorded on the southeast (15–17.5 µg/m³) and southwest (< 15 µg/m³). In addition, with an increase in altitude the concentration of PM_{2.5} decreases. The spatial pattern of PM_{2.5} in Serbia follows the distribution of the population with the highest values for urban areas (> 22.5 µg/m³).
- For most parts of Serbia, there is a negative trend of annual PM_{2.5}, with the highest values in the eastern part of the country (–0.37 µg/m³ per year). This is confirmed by the findings of other studies that are using similar datasets and periods to analyze trends of annual PM_{2.5}. In Europe,

a significant reduction in emissions of gases (that are precursors of secondary PM) since 1990, is mainly related to the changes in energy and industry sectors. In Serbia, for the period 1990–2016, the NO_x, SO_x, NH₃, and NMVOCs showed a slight decrease in emissions.

- Despite the negative trend, the problem of air pollution, i.e., PM_{2.5}, became a leading challenge in most urban environments mainly due to the emission from household heating, traffic, industry, and other urban activities. It is shown that, in Serbia, with the increase of the size of the urban area, the concentration of PM_{2.5} increase as well. The highest observed concentrations are not always recorded for the largest urban areas due to the influence of many factors (local topography, meteorology, and specific of PM sources). High level of pollutants has been recorded in many cities in Serbia during the winter months, due to activation of large number of individual household heating. In 2018, heating plants with a capacity below 50 MW, and individual domestic household heating contribute with 75% in total emissions of PM_{2.5}.
- It is shown that more than 99% of the territory of Serbia has the mean annual values under the national limit established by regulation (25 µm/m³), but comparing to WHO standard (10 µm/m³), the whole territory is above the safe level. In line with urbanization, there is a clear upward trend in the share of the population exposed to higher values of PM_{2.5}. The share of the population exposed to values higher than 25 µg/m³ increased from 6.65% in 2005 to 11.4% in 2015. Comparing to the WHO limit, the total population in Serbia is exposed to higher values in all the observed years (2005, 2010, and 2015).

Despite some advances in spatio-temporal resolution of the existing PM_{2.5} datasets, there are still uncertainties mainly due to the insufficient number of ground-level measurements. The larger spatial coverage by ground-level monitoring stations can improve model accuracy as well data about chemical composition. Also, this is important for understanding the small-scale processes, especially in urban environments where the population is exposed to the highest air pollution. The significant efforts must be performed in the area of public policies to reduce emissions and population exposures. This study aims to contribute to scientific knowledge about PM_{2.5} for the territory of Serbia. Further research will focus on the local predictors of PM_{2.5} distribution and episodes with extremely high values of PM.

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References

- Aničić, M., Spasić, T., Tomašević, M., Rajšić, S., & Tasić, M. (2011). Trace elements accumulation and temporal trends in leaves of urban deciduous trees (*Aesculus hippocastanum* and *Tilia* spp.). *Ecological Indicators*, 11(3), 824–830. <https://doi.org/10.1016/j.ecolind.2010.10.009>
- Atmospheric Composition Analysis Group. (n.d.). *Surface PM2.5: European Regional Estimates (V4.EU.02)* [Data set]. Retrieved from http://fizz.phys.dal.ca/~atmos/martin/?page_id=140
- Bai, K., Ma, M., Chang, N.-B., & Gao, W. (2019). Spatiotemporal trend analysis for fine particulate matter concentrations in China using high-resolution satellite-derived and ground-measured PM_{2.5} data. *Journal of Environmental Management*, 233, 530–542. <https://doi.org/10.1016/j.jenvman.2018.12.071>
- Beelen, R., Hoek, G., van den Brandt, P. A., Goldbohm, R. A., Fischer, P., Schouten, L. J., . . . Brunekreef, B. (2008). Long-Term Effects of Traffic-Related Air Pollution on Mortality in a Dutch Cohort (NLCS-AIR Study). *Environmental Health Perspectives*, 116(2), 196–202. <https://doi.org/10.1289/ehp.10767>

- Boys, B. L., Martin, R. V., van Donkelaar, A., MacDonell, R. J., Hsu, N. C., Cooper, M. J., . . . Wang, S. W. (2014). Fifteen-Year Global Time Series of Satellite-Derived Fine Particulate Matter. *Environmental Science and Technology*, 48(19), 11109–11118. <https://doi.org/10.1021/es502113p>
- Brauer, M., Hoek, G., van Vliet, P., Meliefste, K., Fischer, P., Gehring, U., . . . Brunekreef, B. (2003). Estimating long-term average particulate air pollution concentrations: application of traffic indicators and geographic information systems. *Epidemiology*, 14(2), 228–239. <https://doi.org/10.1097/01.EDE.0000041910.49046.9B>
- Brunekreef, B., & Holgate, S. T. (2002). Air pollution and health. *The Lancet*, 360(9341), 1233–1242. [https://doi.org/10.1016/S0140-6736\(02\)11274-8](https://doi.org/10.1016/S0140-6736(02)11274-8)
- Center for International Earth Science Information Network [CIESIN]. (2018a). *Gridded Population of the World, Version 4 (GPW, v4): Population Count Adjusted to Match 2015 Revision of UN WPP Country Totals, Revision 11*. <https://doi.org/10.7927/H4JW8BX5>
- Center for International Earth Science Information Network [CIESIN]. (2018b). *Gridded Population of the World, Version 4 (GPW, v4): Population Density Adjusted to Match 2015 Revision UN WPP Country Totals, Revision 11*. <https://doi.org/10.7927/H4F47M65>
- Cheng, B., & Wang-Li, L. (2019). Spatial and Temporal Variations of PM_{2.5} in North Carolina. *Aerosol and Air Quality Research*, 19(4), 698–710. <https://doi.org/10.4209/aaqr.2018.03.0111>
- Cohen, A. J., Brauer, M., Burnett, A., Anderson, H. R., Frostad, J., Estep, K., . . . Forouzanfar, M. H. (2017). Estimates and 25-year trends of the global burden of disease attributable to ambient air pollution: an analysis of data from the Global Burden of Diseases Study 2015. *The Lancet*, 389(10082), 1907–1918. [https://doi.org/10.1016/S0140-6736\(17\)30505-6](https://doi.org/10.1016/S0140-6736(17)30505-6)
- Conrad, O., Bechtel, B., Bock, M., Dietrich, H., Fischer, E., Gerlitz, L., . . . Böhner, J. (2015). System for Automated Geoscientific Analyses (SAGA) v. 2.1.4. *Geoscientific Model Development*, 8, 1991–2007. <https://doi.org/10.5194/gmd-8-1991-2015>
- Cvetković, Ž., Logar, M., Rosić, A., & Ćirić, A. (2012). Mineral composition of the airborne particles in the coal dust and fly ash of the Kolubara basin (Serbia). *Periodico di Mineralogia*, 81(2), 205–223. <https://doi.org/10.2451/2012PM0012>
- Djordjevic, P., Nikolic, Dj., Jovanovic, I., Mihajlovic, I., Savic, M., & Zivkovic, Z. (2013). Episodes of extremely high concentrations of SO₂ and particulate matter in the urban environment of Bor, Serbia. *Environmental Research*, 126, 204–207. <https://doi.org/10.1016/j.envres.2013.05.002>
- Drobnjaković, M., & Spalević, A. (2017). Naselja Srbije (The settlements in Serbia). In M. Radovanović (Ed.), *Geografija Srbije* (Geography of Serbia) (pp. 566–613). Belgrade, Serbia: Geografski institute „Jovan Cvijić“ SANU.
- European Environment Agency. (2013). *European Digital Elevation Model (EU-DEM), version 1.1* [Data set]. Retrieved from <https://land.copernicus.eu/imagery-in-situ/eu-dem/eu-dem-v1.1?tab=download>
- European Environment Agency. (2018). *Air quality in Europe — 2018 report*, EEA Report No 12/2018. Retrieved from <https://www.eea.europa.eu/publications/air-quality-in-europe-2018/air-quality-in-europe-2018/viewfile#pdfjs.action=download>
- European Environment Agency. (2019). *Emissions of the main air pollutants in Europe*. Retrieved from <https://www.eea.europa.eu/downloads/87a94e27cde74821b15b48e77e062b69/1567590774/assessment-6.pdf>
- European Parliament, & Council of the European Union. (2008). *Directive 2008/50/EC of the European Parliament and of the Council of 21 May 2008: On Ambient Air Quality and Cleaner Air for Europe*. Retrieved from <https://eur-lex.europa.eu/legal-content/EN/TXT/PDF/?uri=CELEX:32008L0050&from=en>
- European Parliament, & Council of the European Union. (2013). *Decision No 1386/2013/EU of the European Parliament and of the Council of 20 November 2013 on a General Union Environment Action Programme to 2020 'Living well, within the limits of our planet'*. Retrieved from <http://eur-lex.europa.eu/legal-content/EN/TXT/PDF/?uri=CELEX:32013D1386&from=EN>
- Fuzzi, S., Baltensperger, U., Carslaw, K., Decesari, S., Denier van der Gon, H., Facchini, M. C., . . . Gilardoni, S. (2015). Particulate matter, air quality and climate: lessons learned and future needs. *Atmospheric Chemistry and Physics*, 15, 8217–8299. <https://doi.org/10.5194/acp-15-8217-2015>

- Giannadaki, D., Giannakis, E., Pozzer, A., & Lelieveld, J. (2018). Estimating health and economic benefits of reductions in air pollution from agriculture. *Science of the Total Environment*, 622–623, 1304–1316. <https://doi.org/10.1016/j.scitotenv.2017.12.064>
- Hadžić, M., & Zeković, S. (2019). Rethinking deindustrialization, and the reindustrialization policy in Serbia. *Spatium*, 41, 14–22. <https://doi.org/10.2298/SPAT1941014H>
- Joksić, J. D., Jovašević-Stojanović, M., Bartonova, A., Radenković, M. B., Yttri, K.-E., Matić-Besarabić, S., & Ignjatović, Lj. (2009). Physical and chemical characterization of the particulate matter suspended in aerosols from the urban area of Belgrade. *Journal of the Serbian Chemical Society*, 74(11), 1319–1333. <https://doi.org/10.2298/JSC0911319J>
- Jovanović, M. (Ed.). (2019). *Godišnji izveštaj o stanju kvaliteta vazduha u Republici Srbiji 2018. godine* [Annual report on the state of air quality in the Republic of Serbia in 2018]. Retrieved from http://www.sepa.gov.rs/download/izv/Vazduh2018_final.pdf
- Jovašević-Stojanović, M., Bartonova, A., Topalović, D., Lazović, I., Pokrić, B., & Ristovski, Z. (2015). On the use of small and cheaper sensors and devices for indicative citizen-based monitoring of respirable particulate matter. *Environmental Pollution*, 206, 696–704. <https://doi.org/10.1016/j.envpol.2015.08.035>
- Karagulian, F., Belis, C. A., Dora, C. F. C., Prüss-Ustün, A. M., Bonjour, S., Adair-Rohani, H., & Amann, M. (2015). Contributions to cities' ambient particulate matter (PM): A systematic review of local source contributions at global level. *Atmospheric Environment*, 120, 475–483. <http://dx.doi.org/10.1016/j.atmosenv.2015.08.087>
- Kiesewetter, G., & Amann, M. (2014). *Urban PM_{2.5} levels under the EU Clean Air Policy Package* (TSAP Report #12). Retrieved from https://ec.europa.eu/environment/air/pdf/TSAP_12.pdf
- Kokotović Kanazir, V., Stojilković Gnjatović, J., Filipović, M., Babović, S., Ivković, M., & Lović Obradović, S. (2017). Stanovništvo Srbije (Population of Serbia). In M. Radovanović (Ed.) *Geografija Srbije* (Geography of Serbia) (pp. 506–613). Belgrade, Serbia: Geografski institute „Jovan Cvijić“ SANU.
- Kovačević, G. C. (2019). *Procena uticaja kvaliteta vazduha na pogoršanje alergijskog rinitisa i astme* (Doktorska disertacija) [Assessment of the air quality impact on allergic rhinitis and asthma exacerbation (Doctoral dissertation)]. Univerzitet u Beogradu, Medicinski fakultet, Beograd. Retrieved from <http://uvidok.rcub.bg.ac.rs/bitstream/handle/123456789/3322/Doktorat.pdf?sequence=1>
- Krunić Lazić, M. (Ed.). (2018). *Privredni potencijali i aktivnosti od značaja za životnu sredinu Republike Srbije 2017.* [The economic potential and activities of importance to the environment in the Republic of Serbia in 2017]. Retrieved from https://www.sepa.gov.rs/download/posebni/Privredne_aktivnosti_2017.pdf
- Larkin, A., van Donkelaar, A., Geddes, J. A., Martin, R. V., & Hystad, P. (2016). Relationships between Changes in Urban Characteristics and Air Quality in East Asia from 2000 to 2010. *Environmental Science & Technology*, 50(17), 9142–9149. <https://doi.org/10.1021/acs.est.6b02549>
- Lekić, D. (Ed.). (2018). *Izveštaj o stanju životne sredine u Republici Srbiji za 2017. godinu* [The report on the environmental situation in the Republic of Serbia for 2017]. Retrieved from http://www.sepa.gov.rs/download/Izvestaj_2017.pdf
- Li, J., Han, X., Jin, M., Zhang, X., & Wang, S. (2019). Globally analysing spatiotemporal trends of anthropogenic PM_{2.5} concentration and population's PM_{2.5} exposure from 1998 to 2016. *Environment International*, 128, 46–62. <https://doi.org/10.1016/j.envint.2019.04.026>
- Lu, D., Mao, W., Yang, D., Zhao, J., & Xu, J. (2018). Effects of land use and landscape pattern on PM_{2.5} in Yangtze River Delta, China. *Atmospheric Pollution Research*, 9(4), 705–713. <https://doi.org/10.1016/j.apr.2018.01.012>
- Malinović-Miličević, S. B., Mihailović, D. T., Nikolić-Đorić, E. B., & Jevtić, M. R. (2015). Gaseous and particulate urban air pollution in the region of Vojvodina (Serbia). *Matica Srpska Journal of Natural Sciences*, 128, 87–97. <https://doi.org/10.2298/ZMSPN1528087M>
- Perrone, M. G., Vratolis, S., Georgieva, E., Török, S., Šega, K., Veleva, B., . . . Belis, C. A. (2018). Sources and geographic origin of particulate matter in urban areas of the Danube macro-region: The cases of Zagreb (Croatia), Budapest (Hungary) and Sofia (Bulgaria). *Science of the Total Environment*, 619–620, 1515–1529. <https://doi.org/10.1016/j.scitotenv.2017.11.092>
- Pope III, C. A., & Dockery, D. W. (2006). Health Effects of Fine Particulate Air Pollution: Lines that Connect. *Journal of the Air & Waste Management Association*, 56(6), 709–742. <https://doi.org/10.1080/10473289.2006.10464485>

- Querol, X., Alastuey, A., Moreno, T., Viana, M. M., Castillo, S., Pey, J., . . . Sánchez de la Campa, A. (2008). Spatial and temporal variations in airborne particulate matter (PM₁₀ and PM_{2.5}) across Spain 1999–2005. *Atmospheric Environment*, 42(17), 3964–3979. <https://doi.org/10.1016/j.atmosenv.2006.10.071>
- Radović, F., & Jovanović, M. (Eds.). (2017). *Godišnji izveštaj o stanju kvaliteta vazduha u Republici Srbiji 2016. godine* [Annual report on the state of air quality in the Republic of Serbia in 2016]. Retrieved from <http://www.sepa.gov.rs/download/VAZDUH2016.pdf>
- Radović, F., & Jovanović, M. (Eds.). (2018). *Godišnji izveštaj o stanju kvaliteta vazduha u Republici Srbiji 2017. godine* [Annual report on the state of air quality in the Republic of Serbia in 2017]. Retrieved from <http://www.sepa.gov.rs/download/VAZDUH2017.pdf>
- Rajšić, S. F., Tasić, M. D., Novaković, V. T., & Tomašević, M. N. (2004). First Assessment of the PM₁₀ and PM_{2.5} Particulate Level in the Ambient Air of Belgrade City. *Environmental Science and Pollution Research*, 11(3), 158–164. <https://doi.org/10.1007/BF02979670>
- Reizer, M., & Juda-Rezler, K. (2016). Explaining the high PM₁₀ concentrations observed in Polish urban areas. *Air Quality, Atmosphere & Health*, 9(5), 517–531. <https://doi.org/10.1007/s11869-015-0358-z>
- Rogula-Kozłowska, W., Klejnowski, K., Rogula-Kopiec, P., Osródka, L., Krajny, E., Błaszczak, B., & Mathews, B. (2014). Spatial and seasonal variability of the mass concentration and chemical composition of PM_{2.5} in Poland. *Air Quality, Atmosphere & Health*, 7(1), 41–58. <https://doi.org/10.1007/s11869-013-0222-y>
- Saarikoski, S. K., Sillanpää, M. K., Saarnio, K. M., Hillamo, R. E., Pennanen, A. S., & Salonen, R. O. (2008). Impact of Biomass Combustion on Urban Fine Particulate Matter in Central and Northern Europe. *Water, Air, & Soil Pollution*, 191(1–4), 265–277. <https://doi.org/10.1007/s11270-008-9623-1>
- Samek, L., Stegowski, Z., Furman, L., Styszko, K., Szramowiat, K., & Fiedor, J. (2017). Quantitative Assessment of PM_{2.5} Sources and Their Seasonal Variation in Krakow. *Water, Air, & Soil Pollution*, 228(8), 290. <https://doi.org/10.1007/s11270-017-3483-5>
- Shaddick, G., Thomas, M. L., Amini, H., Broday, D., Cohen, A., Frostad, J., . . . Brauer, M. (2018). Data Integration for the Assessment of Population Exposure to Ambient Air Pollution for Global Burden of Disease Assessment. *Environmental Science & Technology*, 52, 9069–9078. <https://doi.org/10.1021/acs.est.8b02864>
- Shisong, C., Wenji, Z., Hongliang, G., Deyong, H., You, M., Wenhui, Z., & Shanshan, L. (2018). Comparison of remotely sensed PM_{2.5} concentrations between developed and developing countries: Results from the US, Europe, China, and India. *Journal of Cleaner Production*, 182, 672–681. <https://doi.org/10.1016/j.jclepro.2018.02.096>
- Stamenković, S., Djekić, T., Ristić, S., Novković, V., Mitrović, T., & Marković, M. (2016). Air quality lichen monitoring at three selected urban areas in the Southern Serbia. *Biologica Nyssana*, 7(1), 19–29. <http://doi.org/10.5281/zenodo.159100>
- Statistical Office of the Republic of Serbia. (2013). *Popis stanovništva, domaćinstava i stanova 2011. u Republici Srbiji: Knjiga 30. Stanovi prema vrsti energenta za grejanje, podaci po opštinama/gradovima* [2011 Census of Population, Households and Dwellings in the Republic of Serbia: Book 30. Dwellings by the type of energy raw material used for heating, data by municipalities/cities]. Retrieved from <https://publikacije.stat.gov.rs/G2013/Pdf/G20134023.pdf>
- Statistical Office of the Republic of Serbia. (2014). *Popis stanovništva, domaćinstava i stanova 2011. u Republici Srbiji: Knjiga 20. Upporedni pregled broja stanovnika 1948, 1953, 1961, 1971, 1981, 1991, 2002. i 2011., podaci po naseljima* [Comparative overview of the number of population in 1948, 1953, 1961, 1971, 1981, 1991, 2002 and 2011, data by settlements]. Retrieved from <http://publikacije.stat.gov.rs/G2014/Pdf/G20144008.pdf>
- Statistical Office of the Republic of Serbia. (2019). *Statistički godišnjak Republike Srbije 2019* [Statistical Yearbook of the Republic of Serbia 2019]. Retrieved from <https://publikacije.stat.gov.rs/G2019/Pdf/G20192052.pdf>
- Stevanović, I., Jovašević-Stojanović, M., & Jović Stošić, J. (2016). Association between ambient air pollution, meteorological conditions and exacerbations of asthma and chronic obstructive pulmonary disease in adult citizens of the town of Smederevo. *Vojnosanitetski pregled*, 73(2), 152–158. <https://doi.org/10.2298/VSP141110265>
- Šerbula, S. M., Antonijević, M. M., Milošević, N. M., Milić, S. M., & Ilić, A. A. (2010). Concentrations of particulate matter and arsenic in Bor (Serbia). *Journal of Hazardous Materials*, 181(1–3), 43–51. <https://doi.org/10.1016/j.jhazmat.2010.04.065>

- Tasic, M., Rajsic, S., Novakovic, V., & Mijic, Z. (2007). An assessment of air quality in Belgrade urban area: PM₁₀, PM_{2.5} and trace metals. *Journal of Physics: Conference Series*, 71, 012016. Retrieved from <https://iopscience.iop.org/article/10.1088/1742-6596/71/1/012016/pdf>
- Tasić, M, Rajšić, S., Novaković, V., & Mijić, Z. (2006). Atmospheric aerosols and their influence on air quality in urban areas. *Facta Universitatis, Series: Physics, Chemistry and Technology*, 4(1), 83–91. Retrieved from <https://pdfs.semanticscholar.org/85fd/694e5872932c74d6260301311f6f72f59308.pdf>
- Tasić, V. (2017). Using low-cost sensors and systems for air pollution monitoring. *Proceedings of 5th International Conference on Renewable Electrical Power Sources*, 5(1), 11–19. <https://doi.org/10.24094/mkoiee.017.5.1.11>
- Tasić, V., Milošević, N., Kovačević, R., Jovašević-Stojanović, M., & Dimitrijević, M. (2012). Indicative levels of PM in the ambient air in the surrounding villages of the copper smelter complex Bor, Serbia. *Chemical Industry & Chemical Engineering Quarterly*, 18(4–2), 643–652, <https://doi.org/10.2298/CICEQ111228109T>
- Todorović, M. N., Radenković, M. B., Rajšić, S. F., & Ignjatović, Lj. M. (2019). Evaluation of mortality attributed to air pollution in the three most populated cities in Serbia. *International Journal of Environmental Science and Technology*, 16(11), 7059–7070. <https://doi.org/10.1007/s13762-019-02384-6>
- United Nations. (2015). *Transforming our world: the 2030 Agenda for Sustainable Development*. Retrieved from https://www.un.org/en/development/desa/population/migration/generalassembly/docs/globalcompact/A_RES_70_1_E.pdf
- Uredba o uslovima za monitoring i zahtevima kvaliteta vazduha [Regulation on Monitoring Conditions and Air Quality Requirements]. Službeni glasnik Republike Srbije br. 11 (2010); 75 (2010); 63 (2013).
- van Donkelaar, A., Martin, R. V., Brauer, M., & Boys, B. L. (2015). Use of Satellite Observations for Long-Term Exposure Assessment of Global Concentrations of Fine Particulate Matter. *Environmental Health Perspectives*, 123(2), 135–143. <https://doi.org/10.1289/ehp.1408646>
- van Donkelaar, A., Martin, R. V., Brauer, M., Hsu, N. C., Kahn, R. A., Levy, R. C., . . . Winker, D. M. (2016). Global Estimates of Fine Particulate Matter using a Combined Geophysical-Statistical Method with Information from Satellites, Models, and Monitors. *Environmental Science & Technology*, 50(7), 3762–3772. <https://doi.org/10.1021/acs.est.5b05833>
- van Donkelaar, A., Martin, R. V., Li, C., & Burnett, R. T. (2019). Regional Estimates of Chemical Composition of Fine Particulate Matter using a Combined Geoscience-Statistical Method with Information from Satellites, Models, and Monitors. *Environmental Science & Technology*, 53(5), 2595–2611. <https://doi.org/10.1021/acs.est.8b06392>
- van Donkelaar, A., Martin, R. V., Brauer, M., Kahn, R., Levy, R., Verduzco, C., & Villeneuve, P. J. (2010). Global Estimates of Ambient Fine Particulate Matter Concentrations from Satellite-Based Aerosol Optical Depth: Development and Application. *Environmental Health Perspectives*, 118(6), 847–855. <https://doi.org/10.1289/ehp.0901623>
- Vicente, E. D., & Alves, C. A. (2018). An overview of particulate emissions from residential biomass combustion. *Atmospheric Research*, 199, 159–185. <https://doi.org/10.1016/j.atmosres.2017.08.027>
- Vodonos, A., Awad, Y. A., & Schwartz, J. (2018). The concentration-response between long-term PM_{2.5} exposure and mortality; A meta-regression approach. *Environmental Research*, 166, 677–689. <https://doi.org/10.1016/j.envres.2018.06.021>
- Vuković, G., Aničić Urošević, M., Pergal, M., Janković, M., Goryainova, Z., Tomašević, M., & Popović, A. (2015). Residential heating contributon to level air pollutants (PAHs, major, trace, and rare earth elements): a moss bag case study. *Environmental Science and Pollution Ressearch*, 22(23), 18956–18966. <https://doi.org/10.1007/s11356-015-5096-0>
- Wang, P., Huang, C., Brown de Colstoun, E. C., Tilton, J. C., & Tan, B. (2017). Global Human Built-up And Settlement Extent (HBASE) Dataset From Landsat. Palisades, NY: NASA Socioeconomic Data and Applications Center (SEDAC). <https://doi.org/10.7927/H4DN4345>
- Wang, Q., Kwan, M.-P., Zhou, K., Fan, J., Wang, Y., & Zhan, D. (2019). The impacts of urbanization on fine particulate matter (PM_{2.5}) concentrations: Empirical evidence from 135 countries worldwide. *Environmental Pollution*, 247, 989–998. <https://doi.org/10.1016/j.envpol.2019.01.086>
- World Health Organization. (2018). *Burden of disease from ambient air pollution for 2016*. Retrieved from https://www.who.int/airpollution/data/AAP_BoD_results_May2018_final.pdf

- World Health Organization Regional Office for Europe. (2017). *Evolution of WHO air quality guidelines: past, present and future*. Retrieved from http://www.euro.who.int/__data/assets/pdf_file/0019/331660/Evolution-air-quality.pdf
- World Health Organization. (2006). *WHO Air quality guidelines for particulate matter, ozone, nitrogen dioxide and sulfur dioxide: Global update 2005 – Summary of Risk Assessment*. Retrieved from https://apps.who.int/iris/bitstream/handle/10665/69477/WHO_SDE_PHE_OEH_06.02_eng.pdf?sequence=1&isAllowed=y
- World Health Organization Regional Office for Europe. (2019). *Health impact of ambient air pollution in Serbia – a Call to action*. Retrieved from http://www.euro.who.int/__data/assets/pdf_file/0020/412742/Health-impact-pollution-Serbia.pdf?ua=1
- Xiao, Q., Ma, Z., Li, S., & Liu, Y. (2015). The Impact of Winter Heating on Air Pollution in China. *PLoS ONE*, 10(1), e0117311. <https://doi.org/10.1371/journal.pone.0117311>
- Yang, D., Ye, C., Wang, X., Lu, D., Xu, J., & Yang, H. (2018). Global distribution and evolvement of urbanization and PM_{2.5} (1998–2015). *Atmospheric Environment*, 182, 171–178. <https://doi.org/10.1016/j.atmosenv.2018.03.053>